Determining Emissions From Landfills And Creating Odor Buffer Distances

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DETERMINING
EMISSIONS FROM LANDFILLS AND
CREATING ODOR BUFFER DISTANCES

by

NICHOLAS SCOTT GUARRIELLO
B.S. Florida State University, 2007

A thesis submitted in partial fulfillment of the requirements
for the degree of Master of Science Environmental Engineering
in the Department of Civil and Environmental Engineering
in the College of Engineering and Computer Science
at the University of Central Florida
Orlando, Florida

Spring Term
2009

Major Professor: C. David Cooper
ABSTRACT

With population growing every year, more and more people are looking for places to live. This can lead to construction of houses near and around landfills. As homes get closer to landfills, the odors these landfills produce become more of a problem, and lead to an increase in odor complaints. Modeling these odors and recommending odor buffer distances will help determine limits on how close to landfills new homes should be allowed. This should help reduce future odor complaints.

To solve this problem one must accurately estimate odorous gas emissions from the landfill. Often odors can be indicated by methane emissions. A new technique using hundreds of ambient VOC concentrations, which are taken from landfills on a quarterly basis, was used to invert and solve the Gaussian dispersion equation for methane emissions. In this technique, Voronoi diagram theory was used to automatically locate numerous point sources for optimal positioning relative to receptors. The newly solved methane emission rates can now be input into a dispersion model, and the resulting methane concentrations used as surrogates for odors around the landfill.

One of the most important steps in the analysis is to determine which model is best to use for odor modeling. There are many considerations that go into this decision, such as how much time it takes to run the model, how accurate the model is, and how easy the model is to use. Two current models CALPUFF and AERMOD were compared. In the modeling, methane was used as a surrogate for the odors. Since landfills handle many different combinations of waste, the type of odor may vary from landfill to landfill. In this test case, H$_2$S was assumed to be the main
contributor to the odor emitted from the landfill, and the $\text{H}_2\text{S}$-to-methane ratio was used to estimate downwind $\text{H}_2\text{S}$ concentrations from the modeled methane concentrations.

Once an air dispersion model is selected, it can be used to model odors and to develop a graphical screening method to show where these odors are most likely to occur and how strong they will be. This can be used to determine how close to a landfill homes can be built without having significant odor impacts bothering these new residents. Also, this tool can be used for improving landfill gas management. Several example scenarios include the possibility of not enough soil cover placed on the waste, leaks from an aging collection system, or cracks in the collection piping created by the settling of waste.
ACKNOWLEDGMENTS

I would like to thank my advisor, Dr. C. David Cooper for his support and advice throughout this effort. I would also like to thank Veronica Figueroa and Dr. Kevin Mackie for all their work and contributions to this project, and the Hinkley Center for Solid and Hazardous Waste Management for the funding of this project.

Lastly but most importantly I would like to thank my parents, Nicholas P. and Jean Guarriello, for bringing me into this world and supporting me in every way possible. Without them I would not be the man I am today.
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CHAPTER 1: INTRODUCTION

The research that is presented in this thesis is a continuation of a three-year project that was undertaken to use the latest dispersion modeling to predict odors near landfills and to help recommend odor buffer distances around landfills. The objectives of this three-year project were split into the three major objectives, one per year: “Year 1) The development a modeling methodology for using AERMOD and CALPUFF to predict odors and appropriate odor buffer distances around landfills, and demonstrate the modeling methodology for one selected landfill in Florida; Year 2) The comparison of AERMOD results to CALPUFF results, to determine which model is better suited for the odor buffer study; and Year 3) The development of a simplified modeling tool for use by solid waste managers.”

This thesis is a continuation and revision of this three year project started by Veronica Figueroa. Revisions to the method used to measure methane emissions, an addition of a sensitivity study of this method, a comparison of CALPUFF and AERMOD, and the development of a modeling tool to show the odors around landfills are all described in this thesis.

Data used in this project came from the Seminole County Landfill. The Seminole County Landfill is a Class 1 municipal solid waste landfill located in the middle of 6000 acres of land and has three phases. Phase 1 is closed, capped, and is 131 feet high. Phase 2 is the active part of the landfill and is currently 73 feet high. Phase 3 is between phases 1 and 2, it will be 270 feet high. All three phases make up 232 acres of the disposal area and only 127 acres of this have been used. This landfill serves over 300,000 residents and receives approximately 810 tons per day of different types of municipal solid waste. The landfill has a gas recovery system, and until recently, they simply burned off the landfill gas. On May 8, 2008 the landfill opened a landfill
gas to energy power plant, which will produce electricity to power more than 6,000 homes every year and is expected to operate for more than 40 years. Figure 1 is an aerial photograph of Seminole County Landfill showing the three phases of the landfill.

Figure 1: Three Phases of the Seminole County Landfill

This landfill like most others around the United States creates what is known as landfill gas. This gas (commonly known as biogas) is the gas that is produced when organic matter biologically breaks down in the absence of oxygen. In landfills this gas is mostly comprised of methane (CH₄) 45-60%, carbon dioxide (CO₂) 40-60%, and a large number of trace chemical species. Methane is also a key contributor to global warming and when comparing it to carbon dioxide it is 25 times more potent as a greenhouse gas than carbon dioxide. That is why it is so important to know how much methane is emitted even if the source is emitting more carbon dioxide. As seen in Table 1, landfills are still the largest contributor of methane in the U.S., and are projected to continue to be the largest contributor of methane in the future.
Table 1: U.S. Baseline Methane Emissions, MMTCE\textsuperscript{4}

<table>
<thead>
<tr>
<th></th>
<th></th>
<th></th>
<th></th>
<th></th>
<th></th>
<th></th>
<th></th>
</tr>
</thead>
<tbody>
<tr>
<td>Landfills</td>
<td>59.3</td>
<td>60.8</td>
<td>56.9</td>
<td>55.5</td>
<td>55.1</td>
<td>52.0</td>
<td>47.6</td>
</tr>
<tr>
<td>Coal Mines</td>
<td>24.0</td>
<td>20.3</td>
<td>21.2</td>
<td>22.3</td>
<td>22.4</td>
<td>22.2</td>
<td>21.3</td>
</tr>
<tr>
<td>Natural Gas</td>
<td>33.1</td>
<td>33.9</td>
<td>35.8</td>
<td>36.5</td>
<td>37.4</td>
<td>38.5</td>
<td>39.8</td>
</tr>
<tr>
<td>Manure Management</td>
<td>7.2</td>
<td>8.5</td>
<td>9.4</td>
<td>9.9</td>
<td>10.5</td>
<td>11.2</td>
<td>11.7</td>
</tr>
<tr>
<td>Enteric Fermentation</td>
<td>35.3</td>
<td>37.2</td>
<td>35.1</td>
<td>35.5</td>
<td>36.0</td>
<td>36.5</td>
<td>37.0</td>
</tr>
<tr>
<td>Other*</td>
<td>17.0</td>
<td>16.7</td>
<td>16.6</td>
<td>16.4</td>
<td>16.2</td>
<td>16.5</td>
<td>16.9</td>
</tr>
<tr>
<td><strong>Total</strong></td>
<td><strong>175.8</strong></td>
<td><strong>177.4</strong></td>
<td><strong>175.0</strong></td>
<td><strong>176.2</strong></td>
<td><strong>177.6</strong></td>
<td><strong>177.0</strong></td>
<td><strong>174.2</strong></td>
</tr>
</tbody>
</table>

* - “Other” sources include fossil fuel combustion, oil production, industrial processes, wastewater treatment, rice production, and biomass burning.

The most recent collection of data showing a detailed breakdown of U.S. sources of methane clearly shows in Figure 2 just how much landfills outweigh other sources when it comes to emitting methane.

![Figure 2: 2004 U.S. Sources of Methane\textsuperscript{4}](image)

Table 1: U.S. Baseline Methane Emissions, MMTCE\textsuperscript{4}
Methane is not what causes the odors that people smell around landfills, even though it can be used as a surrogate when it comes to modeling odors. The gas that is the most common source of odors around landfills is hydrogen sulfide (H₂S), but there are many other odorous gases as shown in Table 2 below.⁵

**Table 2: Ranking of Odorous Trace Chemical Species⁵**

<table>
<thead>
<tr>
<th>Odor Rank</th>
<th>Trace Chemical Species</th>
</tr>
</thead>
<tbody>
<tr>
<td>1</td>
<td>Hydrogen sulfide</td>
</tr>
<tr>
<td>2</td>
<td>Methanethiol</td>
</tr>
<tr>
<td>3</td>
<td>Butanoic acid</td>
</tr>
<tr>
<td>4</td>
<td>Ethanal</td>
</tr>
<tr>
<td>5</td>
<td>Carbon disulfide</td>
</tr>
<tr>
<td>6</td>
<td>Ethyl butanoate</td>
</tr>
<tr>
<td>7</td>
<td>1-propanethiol</td>
</tr>
<tr>
<td>8</td>
<td>Dimethyl disulfide</td>
</tr>
<tr>
<td>9</td>
<td>Ethanethiol</td>
</tr>
<tr>
<td>10</td>
<td>1-pentene</td>
</tr>
</tbody>
</table>

The reason hydrogen sulfide ranks so high is its detection limit is so low compared to some of these other trace chemicals that can also cause odors. That does not mean hydrogen sulfide is always going to be the leading source of odors, it is also a function of the emission rate of each of these chemical species. According to the Agency for Toxic Substances and Disease Registry which is part of the Department of Health and Human Services, hydrogen sulfide smells like rotten eggs and has an odor threshold of 0.5-1 ppb.⁶ Landfills in Massachusetts actually have minimum response action levels of exceedances of hydrogen sulfide where they have to go out and fix the problem if there is an exceedance. The minimum response actions are based upon
either exceeding 15 ppb averaged over an 8 hour period or 30 ppb averaged over a 1 hour period. The actions if one of these limits is exceeded are as follows:

1. Log the detection of any exceedances and contact local health officials and the Department within 4 hours for exceedances of the H₂S Action Level.

2. Investigate and determine the source and extent of the exceedance following the protocols in the appendices. If possible, correct the problem immediately. Implement corrective actions, if necessary, including:
   
   1) Cease acceptance of any material that has the potential to contribute to hydrogen sulfide emissions, on at least a temporary basis.
   
   2) Place additional daily and intermediate cover soils or apply other cover technologies to reduce hydrogen sulfide emissions to ambient air.

3. Implement 24-hour continuous air monitoring for hydrogen sulfide in ambient air, and daily near surface monitoring on the landfill. Conduct additional ambient air monitoring off-site or evaluate need for additional off-site monitoring.

4. In addition, the following actions may be required if directed by MassDEP
   
   1) Install a passive landfill gas control system (passive vents) that can be retrofitted to become an active gas collection and control system (combustion and/or non-combustion technologies).
   
   2) Install an active landfill gas control system with landfill gas treatment (combustion and/or non-combustion technologies).
   
   3) Evaluate the need for the installation of a final cover system with an active landfill gas control system on an expedited schedule. Implement a Community Communications Plan, providing notification to the community and local medical/emergency response personnel that hydrogen sulfide concentrations, if they were to migrate off-site, may create an odor nuisance condition.

5. Conduct additional ambient air monitoring off-site to determine the hydrogen sulfide concentration at receptor locations.
Modeling the dispersion of these odorous compounds is done to predict downwind concentrations as a function of distance from a landfill. If detectable and annoying, odors will cause a complaint. It will vary from landfill to landfill, but there are certain parameters that will influence when and how far from the landfill odors will be detected. Parameters include time of day, wind speed, wind direction, temperature, emission rate, local topography, and one's personal sense of an odor. Every person may have a different point at which they will be able to detect an odor and another point at which that odor becomes a nuisance. The time factor of odors is on the order of minutes if not seconds and this also will vary from person to person. These topics of odors are discussed in more detail in the literature review.

It is required for landfills that have the potential to emit more than 50Mg/year of non-methane volatile organic compounds (NMVOCs) to collect and combust their gas (or, with new technologies, to collect and use the gas to produce energy), under 40 CFR Part 60, Subparts WWW of the 1996 EPA New Source Performance Standards and Emission Guidelines for Municipal Solid Waste Landfills. Flaring or converting the landfill biogas to energy will help reduce emissions of odors and other compounds. To make sure these collection systems are running properly, it is required that the landfill do quarterly surface VOC monitoring. An exceedance of 500 ppm above background is a violation of these regulations and will require additional action.

In this study three different sets of this data were used. Appendix A shows the results of the Seminole County Landfill’s Fourth Quarter 2006, Second Quarter 2007, and Second Quarter 2008 Surface Emissions Monitoring Report. The Second Quarter 2008 actually shows an example of an exceedance of 500 ppm and how it was required to come back at a later date to recheck that area. One of the big accomplishments of this research was using the quarterly
surface VOC measurements to solve for not only the total methane emissions from the landfill, but also to resolve the approximate locations of large emissions from within the landfill. Also included in Appendix A with the reports are figures from each of the quarter’s surface emissions walking surveys. These figures show where the field service surveyor started recording the methane measurements and where they ended. The surveyor periodically wrote down the number of the last point he took if there was an exceedance or the surveyor changed directions. Lines were drawn on the plot plan of the route the surveyor took, and using the numbers of the points the surveyor wrote down on the figure it was possible to place each of the points taken as receptors and find their coordinates. This way the receptors can be used to estimate the methane emissions that are discussed in the study.
CHAPTER 2: LITERATURE REVIEW

The goal of this research was to ultimately produce a way to determine odor buffer distances for any particular landfill. There are many ways people have tried to measure or predict odors around different landfills. Since odors can be a nuisance to the people who live around landfills it is important to accurately determine the distance where these odors will first become a problem. There are three main things that lead up to this goal and need to be understood to predict odor buffer distances.

The first of these three main focus points is to understand the fundamental nature of odors, how they are perceived by people, and the different ways they are measured. The next main topic is how odors vary from landfill to landfill. The most common cause of odor from a landfill is H₂S, but in any given landfill, a different compound or group of compounds might be the cause of odors. It is often helpful to determine a ratio of specific odor to methane so that the ultimate goal of accurately determining the odor buffer distance can be achieved. With the ratio of specific odor to methane, one can apply it to the methane concentrations around the landfill and determine the odorous concentrations. The last point that must be discussed is which air dispersion model should be used to model odors. The three models that were compared in this thesis are CALPUFF, AERMOD, and ISC. Each has its benefits, but one model must be chosen, based on accuracy, efficiency, and suitability for modeling odors.

All these topics have been researched by numerous investigators in the past. This author read a number of papers to gather information that was used in this research. The main points of this literature are summarized and presented in the following pages, organized by topic.
The Nature of Odors

An odor can be described as either pleasant or unpleasant, and odors from a landfill fall under the unpleasant category. The actual smell of an odor gives very little information on the compounds in that odor, since there are various compounds that can be mixed together that cause the odor from any landfill as was seen in Table 2. Sulfides and ammonia are the most common odor causing sources in a landfill, but sulfides can cause the strongest of smells since humans can detect them at very low concentrations. Odors are highly unlikely to cause any major health effects, but have been known to cause eye irritations and headaches at high concentrations. It depends on the individual, since every person reacts differently to certain odors. So trying to determine how much odor is coming from a landfill and where it is going can be a very difficult task since wastes have these vastly different chemical compositions.

There are various methods for measuring odorous compounds, each with their advantages and disadvantages. Some examples of these methods are electronic noses (sensor arrays), dynamic dilution olfactometry (using a human panel), using an odor index, and gas chromatographic analysis. A methodology similar to that used in this research is discussed in a study by Tagaris et al. (2003). In that study, CH₄, even though an odorless gas, was used as an index to determine the dispersion of low-reactivity odorous species around a landfill site.⁸ The study used a different method to determine the methane emissions, but did use methane as a surrogate to the odorous chemical species and then used an air dispersion model to track the odors.

Electronic noses contain an array of sensors (sintered metal oxides, catalytic metals, conducting polymers, lipid layers, phtholocyanins, organic semi-conductors, surface acoustic wave or combinations) which respond to a wide variety of chemical classes. The sensors are
based on conducting composites where electrical resistance will change on exposure to a particular vapor. Based on this change in resistance the sensors can identify the type and quantity of the odor. An advantage of the electronic nose is that it can measure a complex group of substances very rapidly. Some disadvantages are it must be standardized by both chemical and olfactometric methods and one of the biggest challenges is detecting complex odors against an intricate background matrix.

Dynamic dilution olfactometry is a way to measure odors that uses humans as the sensors. The most common term used in this method is “odor threshold” or the lowest concentration that can be detected by 50% of the population. The problem with this though is determining the detection of unknown complex mixtures, such as what might be in the air that is around a landfill. A way around this problem is to express the odor strength as a number of odor units, which is a calculated value based on the threshold dilution ratio (human panel). Odor units are commonly used around the U.S. as a way to measure odor emissions, but also have their disadvantages. The main problems with using the odor unit as a standard are the variability of people who serve as the panel that determines the odor units, and the odor unit method includes no measure of importance of the odor. This variability of people who serve as the detectors can be seen in Figure 3. Every person has a different threshold value to where they can detect the odor and another different value for when that odor becomes a nuisance to them.
Dilutions

<table>
<thead>
<tr>
<th>Judge</th>
<th>Dilution Factors (concentrations increase $\rightarrow$)</th>
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<tbody>
<tr>
<td></td>
<td>8 7 6 5 4 3 2 1</td>
</tr>
<tr>
<td></td>
<td>256 128 64 32 16 8 4 2</td>
</tr>
<tr>
<td>1</td>
<td>0 0 0 + + + + +</td>
</tr>
<tr>
<td>2</td>
<td>+ 0 + 0 + + + +</td>
</tr>
<tr>
<td>3</td>
<td>0 + 0 0 0 + + + +</td>
</tr>
<tr>
<td>4</td>
<td>0 0 0 + + + + +</td>
</tr>
<tr>
<td>5</td>
<td>0 0 0 0 0 + + + +</td>
</tr>
<tr>
<td>6</td>
<td>0 0 + + + + + +</td>
</tr>
</tbody>
</table>

"0" indicated that judge selected the wrong sample from a set of three
"+" indicates that judge selected the correct (different) sample

**Figure 3: Calculation of Odor Threshold (Determined by Olfactometry)**

The application of gas chromatography has also been widely used as another way to measure odors. The way this application works is compounds based on their vapor pressures and polarities are separated from the total mixture of the volatile substance. The compounds are then detected as peaks which have specific retention times and peak areas which can be used for qualitative and quantitative determinations, respectively. The major advantage of gas chromatography is that it can separate the different compounds from the air as seen in Figure 4 and determine what type of compounds they are. The major disadvantages include having to extract the air into an adsorbent trap, moisture problems with these traps, and how expensive it is to do this entire process.
This research relied heavily on estimating methane emissions to use as surrogates for odors and there are three main reasons why.

1) Methane represents 40-60% of total emissions from a landfill.

2) The VOC measurements taken every quarter are given as methane (ppm).

3) Methane has a far lower background concentration than carbon dioxide.

There are many ways of estimating methane emissions, each with their advantages and disadvantages just like measuring odors. Some of the ways to estimate methane emissions are...
using flux chambers, biogas production models, optical remote sensing methods, and the method discussed in this thesis using the Gaussian dispersion equation.

The flux chamber has been a popular method used to measure methane emissions from landfills in many studies. There are two ways a flux chamber can be used to measure emissions, either the static/closed flux chamber method or the dynamic/open flux chamber method. The closed flux chamber is the most popular since it is the simple and cheaper to use, but people have discovered errors in its calculations. The major advantage is the closed flux chamber is not being continually diluted with external air so even small fluxes can be measured. The main drawback of the closed chamber technique is as the gas accumulates in the chamber the increase in concentration decreases the flux rate, so the technique underestimates the gas fluxes. The main drawback of the open container is the exact opposite of the closed chamber, it overestimates fluxes due to pressure changes in the flux chamber. Not only will results vary depending on which chamber is used, but other problems that arise with flux chambers are how much time and labor it takes to get one point measurement.

The second method to calculate methane emissions are from biogas production models. The most commonly used biogas production model used is the Landfill Gas Emissions Model (LandGEM) developed by Environmental Protection Agency (EPA). LandGEM is an automated estimation tool with a Microsoft Excel interface that can be used to estimate emission rates for total landfill gas, methane, carbon dioxide, nonmethane organic compounds, and individual air pollutants from municipal solid waste landfills.
\[ Q_{CH_4} = \sum_{i=1}^{n} \sum_{j=0.1}^{1} kL_o \left( \frac{M_i}{10} \right) e^{-kt_{ij}} \]

Where:
- \( Q_{CH_4} \) = annual methane generation in the year of the calculation (m\(^3\)/year)
- \( i = 1 \) year time increment
- \( n = (\text{year of the calculation}) - (\text{initial year of waste acceptance}) \)
- \( j = 0.1 \) year time increment
- \( k \) = methane generation rate (year\(^{-1}\))
- \( L_o \) = potential methane generation capacity (m\(^3\)/Mg)
- \( M_i \) = mass of waste accepted in the \( i^{th} \) year (Mg)
- \( t_{ij} \) = age of the \( j^{th} \) section of waste mass \( M_i \) accepted in the \( i^{th} \) year (decimal years, e.g., 3.2 years)

**Figure 5: First-Order Decomposition Rate Equation\(^{11}\)**

This gas generation model relies heavily on these variables and the assumptions made for variables. Even though this tool is relatively easy to use, the problems that arise with it are that the assumptions will influence the accuracy of the model, and the model is not accounting for how much methane is being emitted versus captured.

Another way methane emissions can be measured is using an optical remote sensing method. The optical remote sensing method uses open-path Fourier transform infrared (OP-FTIR) spectroscopy to obtain path-integrated pollution concentration information along multiple plane-configured optical paths.\(^{12}\) This method has several advantages that some of the other methods previously mentioned don’t have. The measurement of gases is the path-average concentration of each compound within the defined path and it would take many samples along that path for a traditional point sampling method to provide the same data.\(^{13}\) There is no collecting of samples required and sending them back to the lab, which has been a source of errors for other methods. The major disadvantages with this method are the cost, labor to set up, time, and the variability of wind orientation.
Introducing CALPUFF, AERMOD, and ISC

The three most widely used air dispersion models are CALPUFF, AERMOD, and ISC. CALPUFF is a Lagrangian puff model that can predict air quality concentrations over a range of averaging periods and EPA currently approves the CALPUFF model for use for long-range transport of pollutants. AERMOD is a Gaussian plume model that is replacing ISC and is used for short-range transport of pollutants. ISC is similar to AERMOD in that it is also a Gaussian plume model, but is far less advanced and has fewer model options. Also, ISC handles terrain, calm winds, and meteorology differently then AERMOD. There are many studies done comparing these models for all types of pollutants, but not necessarily for odors. That is why it was important to compare each model to see what changes in the models will affect modeling odors.

Starting with a comparison of AERMOD and ISC, a sensitivity study between the two models and their inputs indicates which model is better for modeling odors. Differences found during sensitivity studies between the two models, that might affect modeling odors, are that ISC surface roughness can only be specified as rural or urban, the values in AERMOD can be specified precisely. The mixing heights are calculated by AERMET in AERMOD, but must be specified by the user in ISC. The main difference that affects odor concentrations and showed that ISC was not the best choice for modeling odors is how ISC handled winds. Maximum concentrations predicted by AERMOD and ISC correlated well when wind speeds exceeded 5 m/sec but diverged rapidly as wind speeds decreased. This is caused by the fact that AERMOD includes the effect of plume meander at low wind speeds while ISC does not.
When comparing CALPUFF and AERMOD, each had their advantages and disadvantages. CALPUFF can be used on distances from tens of meters from a source to hundreds of kilometers. It includes algorithms such as building downwash, transitional buoyant and momentum plume rise, partial plume penetration, sub-grid scale terrain, coastal interaction effects, and chemical transformation. Even though CALPUFF has these advantages and algorithms, they are perhaps not useful for short range modeling, such as with odors. CALPUFF has its disadvantages including how much longer it takes to run than AERMOD, the increased amount of inputs CALPUFF has that take time to learn, and understanding what each input does. These can be seen in Table 3.

<table>
<thead>
<tr>
<th>AERMOD</th>
<th>CALPUFF</th>
</tr>
</thead>
<tbody>
<tr>
<td>1. Faster run time (on order of minutes).</td>
<td>1. Long run time (on order of hours).</td>
</tr>
<tr>
<td>2. Fewer inputs to go through.</td>
<td>2. More inputs to go through that are not necessarily used for modeling odors.</td>
</tr>
<tr>
<td>3. Easier to set up an input file and in less time.</td>
<td>3. Longer time to learn model and set up an input file.</td>
</tr>
<tr>
<td>4. Model does a better job a pin-pointing where errors are and what they are.</td>
<td>4. Can take awhile to find errors since the model input file is so large and error file can be difficult to decipher.</td>
</tr>
<tr>
<td>6. Does not track the contribution of plumes from previous hours.</td>
<td>6. Tracks puffs until they have left the modeling domain.</td>
</tr>
</tbody>
</table>

Both CALPUFF and AERMOD have options that can be helpful when modeling, but the model that should be chosen is the one that best models odors. A test case was run for AERMOD and CALPUFF comparing the two models predicted output concentrations to determine which model is the best to use for modeling odors. Both models produced similar outputs for concentrations of odors in the test case, with CALPUFF concentrations being slightly higher. Without measured
odor concentration data to compare with the modeled concentration data, it was a hard decision choosing a model just based on this test. That is why using knowledge of odors, the literature review of the models, the advantages and the disadvantages of each model, and the test case of the models, AERMOD was chosen.
CHAPTER 3: METHODOLOGY

Past researchers have used many of the different methods discussed in the literature review of odors trying to determine the concentrations of odors surrounding landfills. This study had a similar approach to the Targaris methodology in that it used methane as an index for odors and used a ratio of trace chemical species to methane to determine the odor concentrations. Using the methodology shown below, establishing odor buffer distances can be done for any landfill in Florida. A further discussion of each step introduced below is provided in the following pages.

The first task is obtaining ambient air methane concentrations and atmospheric conditions for the landfill to be modeled. The methane concentrations can be obtained from one’s own sampling or from sampling reports that are made every quarter at many Florida landfills. If not measured on site at the time of sampling, some expertise and judgment must be used in estimating the atmospheric conditions (wind speed, wind direction, and stability class) that were present at the time of sampling. These meteorological parameters are needed to complete the task of estimating methane emissions.

The second task is to estimate methane emissions. This task uses an inverse modeling method that is based on the Gaussian dispersion equation. Rearranging the Gaussian dispersion equation and solving the inverse problem, one can predict emission rates at many point locations within the landfill.

The third task is to derive an odorous gas to methane ratio. For this one can use the methane concentrations recorded in task one and a sample of the odorous gas at this same point to determine an average odorous gas to methane ratio for the landfill being modeled.
The limits for acceptable concentrations of this odorous gas are determined in task four and consist of three “threat levels”. The analogy of a stop light is used for these limits and consists of a green, yellow, and red level. As with a stop light green means go or in this case that it is safe to build, and 99.95% of the time odors will not be detected in this area. The color yellow will represent a proceed with caution zone, where depending on the individuals odor threshold odors may be detected from time to time. The third level of red represents a stop and do not build area where odors will be detected and may become a nuisance. This conservative approach is based on using the 4th highest odor concentrations that are determined by dispersion modeling (task five).

The fifth task is to conduct air dispersion modeling using AERMOD to determine where outside the landfill fence-line these odors are going to occur and at what concentrations. For a years worth of meteorological data the 4th highest concentrations of methane are recorded and used to generate color plots of the odor “threat levels.”

Task six involves using the limits set in task four and 4th highest concentrations modeled in task five. Using the limits of methane concentrations which correspond to the odorous gas concentrations coming from the landfill colored plots can be created and one can see the odor buffer distances from an overhead view of the landfill and surrounding areas. Table 4 is a summary of the methodology described above and was used in this study.
<table>
<thead>
<tr>
<th>Task</th>
<th>Description</th>
</tr>
</thead>
<tbody>
<tr>
<td>Task 1</td>
<td><strong>Obtain Methane Concentrations and Atmospheric Conditions</strong> Either by sampling or from existing reports obtain methane concentration data and estimate atmospheric conditions (wind speed, wind direction, and stability class).</td>
</tr>
<tr>
<td>Task 2</td>
<td><strong>Estimate Methane Emissions</strong> Estimate methane emissions using the inverse modeling method.</td>
</tr>
<tr>
<td>Task 3</td>
<td><strong>Derive Odorous Gas to Methane Ratio</strong> Sample landfill gas to estimate odorous gas content and derive odorous gas/methane ratio.</td>
</tr>
<tr>
<td>Task 4</td>
<td><strong>Set Limits</strong> Use ratio to calculate projected methane concentration limits corresponding to red, yellow, and green “threat levels”. (Note suggested limits for hydrogen sulfide are 1-hour ( \text{H}_2\text{S} ) concentrations of &gt;30 ppb for red, 15-30 ppb for yellow, and any concentrations under 15 ppb for green.)</td>
</tr>
<tr>
<td>Task 5</td>
<td><strong>Conduct Air Dispersion Modeling</strong> Gather meteorological data and run AERMOD to determine the distances where these maximum concentrations will occur. (Note-4th highest concentrations represent the 99.95 percentile).</td>
</tr>
<tr>
<td>Task 6</td>
<td><strong>Generate Plots</strong> Using limits set in task 4 and results from air dispersion modeling generate colored plots establishing odor buffer distances.</td>
</tr>
</tbody>
</table>
Task 1: Obtain Methane Concentrations and Atmospheric Conditions

Since many landfills have to do quarterly surface VOC monitoring, such reports were used to estimate methane emissions. VOC monitoring uses numerous measurements, usually anywhere between 350-450 measurements. Using these reports not only provides a robust amount of methane concentrations in the landfill to use to calculate methane emissions, but also helps ensure that no large methane concentrations are missed. If there is a large methane concentration recorded this could reveal a leak in the biogas collection system or a part of the landfill that has developed a crack in the covering. This would not only reveal where a problem is occurring, but also could be a major source of where odors may be coming from.

There is tremendous amount of data given in every quarterly surface VOC monitoring report that can be used in the task of estimating methane concentrations. Not only does each report give the concentrations of methane at each point recorded in the landfill, but gives the upwind and downwind concentrations outside of the landfill. The upwind concentration can be used as the background concentration of methane and is subtracted from the concentrations recorded before being used to estimate methane emissions.

The data in the VOC monitoring reports can be inspected for exceedances (any measurement of methane over 500 ppm). Of the three quarters of reports used in this research, only one quarter (Second Quarter 2008) had exceedances. In the Second Quarter 2008 there were five exceedances of 1052 ppm, 823 ppm, 774 ppm, 1034 ppm, and 559 ppm. Each one was rechecked to make sure there were no problems with the biogas collection system or some other type of leak. The field service surveyor treats each exceedance the same no matter what the concentration is, just as long as it exceeds 500 ppm. First it is rechecked 6-7 days after the initial
exceedance is recorded and then rechecked one month from initial exceedance. Each point at which the exceedance was recorded has to pass both these rechecks, and in the Second Quarter 2008 case, each did.

The next bit of useful data from the quarterly reports is the date and time of each recorded measurement. These were used to help determine the atmospheric conditions at the time of the measurements. It takes about five hours to walk the entire landfill and get between 350-450 measurements. During this time the atmospheric conditions can change, it is best to have fairly constant weather conditions during the walking survey. For this research all three quarters used had relatively constant weather conditions.

Most of the time every quarterly surface VOC monitoring report will come with a NSPS Surface Emissions Monitoring Calibration and Pertinent Data Form. This form is used to retrieve atmospheric conditions that can be used when it comes time to estimate methane emissions. The form has weather observations such as wind speed, wind direction, barometric pressure, air temperature, and other general weather conditions. All these observations are taken by the field service surveyor and are his or her best estimate of the conditions experienced during that time at the landfill.

If the NSPS Surface Emissions Monitoring Calibration and Pertinent Data Form was not attached, there was another way to estimate the atmospheric conditions. The National Weather Service (NWS) archives all the daily measurements of atmospheric conditions recorded at every official NWS data station. There are numerous websites that collect this data and make it available to the public. The website Weather Underground has every atmospheric condition recorded by every hour of the day and sometimes approximately every 10 minutes depending on the weather conditions at the time. Figure 6 below shows examples from the website showing
measurements of atmospheric conditions (by hour for the date of the survey) needed to conduct our modeling. It is noted that the walking survey occurred from about 9 a.m. to 3 p.m. for each of the surveys. These atmospheric conditions and the methane concentrations will be used in the next task of estimating methane emissions.

![Figure 6: Atmospheric Conditions for June 26, 2008](image)

**Task 2: Estimate Methane Emissions**

There are many ways to estimate methane emissions from landfills. Most of these ways were discussed in the literature review such as flux chambers, biogas production models, and
optical remote sensing methods. The method discussed in this thesis is based on the Gaussian dispersion equation, which is shown below as equation (1). This equation (which models the dispersion of a nonreactive gaseous pollutant from an elevated point source) is given here in a form that predicts the steady state concentration (C) in µg/m³ at a point (x, y, z) located downwind from the source.\textsuperscript{18}

\[
C = \frac{Q}{2\pi u \sigma_y \sigma_z} \exp\left(-\frac{1}{2} \frac{y^2}{\sigma_y^2}\right) \exp\left(-\frac{1}{2} \frac{(z-H)^2}{\sigma_z^2}\right) + \exp\left(-\frac{1}{2} \frac{(z+H)^2}{\sigma_z^2}\right)
\]

In this equation \(Q\) is the emission rate (µg/s), \(\sigma_y\) and \(\sigma_z\) are the horizontal and vertical spread parameters (m) (\(\sigma_y\) and \(\sigma_z\) are functions of distance, x, and atmospheric stability), \(u\) is the average wind speed at stack height (m/s), \(y\) is the horizontal distance from the plume centerline (m), \(z\) is the vertical distance above the ground (m), and \(H\) is the effective stack height (\(H= h+\Delta h\), where \(h=\) physical stack height and \(\Delta h=\) plume rise, m).

Since methane emissions from landfills are ground-level sources, \(z = 0\) and \(H = 0\) in equation (1). So equation (1) can be reduced to equation (2) shown below.

\[
C = \frac{Q}{\pi u \sigma_y \sigma_z} \exp\left(-\frac{1}{2} \frac{y^2}{\sigma_y^2}\right)
\]

The equation (2) can be rearranged as equation (3) to solve for \(Q\) the emission rate (for a single source-receptor pair, which will be the methane emission rate needed in this research).

\[
Q = \frac{C}{\exp\left(-\frac{1}{2} \frac{y^2}{\sigma_y^2}\right)}
\]
The horizontal and vertical spread parameters are given as equations (4) and (5), developed by Martin.\textsuperscript{19} The parameters $a$, $b$, $c$, $d$, and $f$ are constants that depend on the stability class and on the distance $x$ ($x$ must be expressed in km).\textsuperscript{18}

\begin{equation}
\sigma_y = ax^b \tag{4}
\end{equation}

\begin{equation}
\sigma_z = cx^d + f \tag{5}
\end{equation}

Equation (3) is the form of the Gaussian dispersion equation that is the foundation for the method presented here for estimating landfill methane emissions. This method closely follows the method described in Figueroa’s thesis, with a few additions and changes to the method. The changes are described in the following paragraphs, but the basic methodology for the matrix inversions used to solve for methane emissions can be found in Figueroa.\textsuperscript{1}

The equations (6) and (7) were used as described in Figueroa’s thesis, but methane background concentration was no longer assumed to be zero.

\begin{equation}
C_{i,j} = f(x, y)_{i,j} Q_j \tag{6}
\end{equation}

\begin{equation}
C_{i,\text{modeled}} = \sum_{j=1}^{n} C_{i,j} \tag{7}
\end{equation}

The upwind methane concentration data from the quarterly surface VOC monitoring report was used as the background concentration and subtracted from each of the recorded methane concentrations ($C_{i,\text{modeled}}$) prior to beginning the solution.
A significant improvement was that the user no longer has to specify (based on judgment and expertise) where to place the sources ($Q_j$) upwind from the receptors ($C_i$). Instead, point source maximum likelihood locations are predicted using Voronoi diagrams, and importance sampling is performed to further refine locations.\textsuperscript{20}

The Voronoi diagram consists of a set of Voronoi regions that can be constructed from a Delaunay tessellation (triangulation) of the existing concentration locations. The vertices of the Voronoi polygons were used to place each source. It is noted that for the case of a landfill, a boundary needs to be set up around the landfill to keep fictitious sources from being created outside the landfill and to prevent the creation of Voronoi polygons with infinite edges. Once the boundary is set the Delaunay triangulation is performed on the receptors inside the landfill, creating a placement of each source relative to the receptors around it. The way each of these sources was placed can be seen in Figure 7 below.

![Figure 7: Example Delaunay Tessellation and Voronoi Regions\textsuperscript{20}](image)
The easiest way to describe how the Delaunay method works is that it forms multiple, non-overlapping triangles by using the three receptors that are closest to each other. For each triangle, the mid point of each line of that triangle was determined and a perpendicular line was drawn from that point. Where three of the perpendicular lines intersected, a point was created and this was considered the source (and a vertex of a corresponding Voronoi polygon). While the sources were being created from the Voronoi diagrams, the sources and receptors were placed to ensure that no source and receptor were within 3m of each other. Once the Voronoi diagram was complete, the sources and the receptors were used to determine the methane emission at each one of the sources.

Importance sampling can also be used to refine the locations and in a paper written by Mackie\textsuperscript{20} shows promise in what importance sampling may be able to do in getting the predicted numbers closer to the actual. Importance sampling is used to determine an optimal placement of source locations by sampling within probability density functions with maximum likelihoods located at the Voronoi vertices.

There are two different methods of importance sampling that show the greatest promise. First is the unconstrained simulation, which moves sources ever so slightly, trying to find better placements while ignoring the solution of emissions found in the non-negative least-squares (NNLS) Voronoi method, used in this thesis. Second is the constrained simulation, which uses the total emissions solved for in the NNLS Voronoi method as a basis for moving the sources. The use of importance sampling is something that will be used in the future for this research, but sensitivity studies on the calculated total emissions using the NNLS Voronoi method described in this section need to be analyzed first.
**Task 3: Derive Odorous Gas to Methane Ratio**

The optimal way to derive the odorous gas to methane ratio would be to sample the landfill gas, estimate the odorous gas content, and then compare it to the methane concentration. That was not possible to do for this research, but would be recommended in further studies. The methodology for determining the odorous gas to methane ratio in this study was done using literature review and doing an average H₂S to methane ratio of similar landfills.

**Task 4: Set Limits**

The ratio calculated from task 3 was used to determine the projected methane concentration limits. The equation (8) below can be used to do this once a ratio is determined or equation (9) can be used if you have the exact percentage of methane and H₂S in the landfill.

\[
\text{Methane Limit (ppm)} = \frac{1}{\left(\frac{H_2S}{CH_4} \times \frac{1000}{(H_2S \text{ Limit (ppb)})}\right)}
\]  

(8)

\[
\text{Methane Limit (ppm)} = \frac{\%CH_4}{\%H_2S} \times \frac{(H_2S \text{ Limit (ppb))}}{1000}
\]

(9)

Note suggested limits are 1-hour H₂S concentrations of >30 ppb for red, 15-30 ppb for yellow, and concentrations under 15 ppb for green (more on these limits is discussed in task 6 of the methodology and in the findings section of this thesis under task 4). To better understand this see example 1 below.

**Example 1**: A landfill contains 53.283 percent by volume methane and 0.002 percent by volume hydrogen sulfide. A contractor wants to build houses near the landfill, but does not want the
hydrogen sulfide concentration of 30ppb to be experienced by the people that will move into these houses. What is the maximum concentration of methane that would be allowed?

**Answer:** Use equation 9

\[
\frac{53.283}{0.002} \times \frac{30 \text{ ppb}}{1000} = \text{Methane Limit (ppm)} = 799 \text{ ppm}
\]

Using equation (8) the suggested limits for H₂S concentrations of >30 ppb, 15-30 ppb, and under 15 ppb can be used to determine the methane concentration limits of any landfill, if the H₂S to methane ratio is known. This was used to generate the plots for odor buffer distances in task 6.

**Task 5: Conduct Air Dispersion Modeling**

AERMOD was determined to be best for the air dispersion modeling of odors. AERMOD stands for American Meteorological Society/Environmental Protection Agency Regulatory Model and as of December 2006 it replaced ISCST3 as the EPA-preferred regulatory model. The model is intended to be used for transport distances up to 50 km and accounts for complex terrain, flat terrain, elevated sources, surface sources, point sources, area sources, calm winds, planetary boundary effects, and temperature just to name a few.

AERMOD has two pre-processors AERMET and AERMAP, which must be run before AERMOD. AERMET is used as the meteorological processor, and consists of NWS surface data, upper air data, and on-site data. AERMET needs both meteorological inputs and surface characteristic inputs to create a surface and profile file that will be fed into AERMOD.

AERMAP is the terrain processor and produces terrain base elevations for each receptor and source. It needs standardized computer files of terrain data to be able to run, which is
available in three formats. The format used in this research was the DEM format (Digital Elevation Model).

**Figure 8: AERMOD Flow Diagram**

When both AERMET and AERMAP were run and the source and receptor information was entered, AERMOD was ready to be used. The user then ran AERMOD for a specified time period of meteorological data, which can be anywhere from a certain hour of a year, to a day of a year, or even an entire year. In this case, one entire year’s worth of meteorological data was used. The output concentrations from AERMOD were recorded at 300 receptors, and then used in the last task to generate plots of these output concentrations.

The 4th highest ranked concentrations were used to generate the plots. The 4th highest is out of 8760 data points, at each receptor, and thus represent the 99.95th percentile. These were used as a way of being extra conservative in predicting odor buffer distances. Other percentiles could be selected; lower ones would lead to shorter distances.
Task 6: Generate Plots

The limits set in task 4 and results from air dispersion modeling were used in establishing odor buffer distances. As an example, Figure 9 shows how odor buffer distances can be established. (The meanings of the colors are explained below)

Figure 9: Example of Odor Buffer Distances

In the example in Figure 9 the distances (x-y scale of graph is in meters) were set into three different colors of red, yellow, and green. The colors represent suggested areas around the landfill where odors may occur. The color red (stop) corresponds to an area where there is the risk of exceeding 30 ppb H₂S, yellow (proceed with caution) corresponds to odors in the range of
15-30 ppb, and green corresponds to the safe zone (where houses may be built and very rarely will an odor from the landfill ever be a nuisance). This is just an example and the methane concentrations on this figure are just showing what is possible. In specific cases, the methane concentrations correspond to the ratio calculated for that particular landfill and for the meteorology around that landfill.

When this is done for any particular landfill more than one plot should be created since methane emissions will change from quarter to quarter and year to year. Making more than one plot can help in creating more of a certainty of where the odor buffer distances will in fact need to be set to ensure odor complaints are reduced to a minimum. Also running the dispersion model for another year’s worth of meteorology and generating a plot is recommended to increase confidence in the odor buffer distances.
CHAPTER 4: FINDINGS

This entire chapter applies the methodology previously discussed to establish odor buffer distances around the Seminole County Landfill (SCL).

Task 1: Estimate Methane Emissions using Three Quarters of SCL data

The three cases of quarterly surface emissions monitoring reports used from the Seminole county landfill consist of 4th Quarter 2006, 2nd Quarter 2007, and 2nd Quarter 2008. Copies of the three reports can be found in Appendix A. Also attached in Appendix A are the NSPS Surface Emissions Monitoring Calibration and Pertinent Data Forms (if one was provided), a history of the weather that occurred on that day, and a figure showing the VOC reading locations.

The quarterly surface emissions monitoring report from 4th Quarter 2006 was taken on December 22, 2006. A total of 425 points were surveyed for VOC concentrations as methane in ppm, using a Landtec SEM 500 flame ionization detector. Based on the NSPS Surface Emissions Monitoring Calibration and Pertinent Data Form, and history from a local weather station, the local wind speed was about 2 mph or 0.89 m/sec, the wind direction was from 130° (SE), and the temperature was about 70°F with mostly clear skies. Based on the information given, the stability class was estimated as class B stability conditions.

The quarterly surface emissions monitoring report from 2nd Quarter 2007 was taken on June 29, 2007. A total of 358 points were surveyed for VOC concentrations as methane in ppm, using a Landtec SEM 500 flame ionization detector. Based on the NSPS Surface Emissions Monitoring Calibration and Pertinent Data Form, and history from a local weather station, the local wind speed was about 3 mph or 1.34 m/sec, the wind direction was from 40° (NE), and the
temperature was about 88°F with scattered clouds. Based on information given, the stability class was estimated as class B stability conditions.

Figure 10: Placement of the Sources and Receptors

The quarterly surface emissions monitoring report from 2nd Quarter 2008 was taken on June 26, 2008. A total of 442 points were surveyed for VOC concentrations as methane in ppm, using a Foxboro TVA-1000B flame ionization detector. Based on the history from a local weather station, the local wind speed was about 5 mph or 2.24 m/sec, the wind direction was from 150° (SE), and the temperature was about 80°F with mostly clear skies. Based on the information given, the stability class was estimated as class B stability conditions. The figure above is an example from the 2nd Quarter 2008 data showing where the wind was blowing toward in the upper right corner, the placement of the receptors, and the placement of the sources which was described in task 2.
Task 2: Estimate Methane Emissions

To estimate where the methane emissions are coming from, and how much each source is emitting, the methodology was used as discussed previously. Using the figures showing the VOC reading locations (located in Appendix A) from task 1 for each quarter, an Excel file was set up with each receptor location and the concentration recorded at that location. Each of the three data sets that show how the input concentration files were set up can be found in Appendix B. Once the concentration file was made, the only thing left to do was change the (run_cases.m) file to match the quarter that was about to be run. In this file, the only parameters that were changed were the wind angle, wind speed, and stability class according to the quarter given. Now the inverse modeling method was run through MATLAB to get the source strengths, and the Voronoi diagram technique was used to place the sources. The source locations (in UTM (km) coordinates) and their emissions (in (ug/s)) were sent to an output file called temp.out. This file was put in excel, where emissions were converted to g/s and summed up to give the total emissions of the landfill for that quarter. Each excel sheet set up from the output files are located in Appendix C for each quarter.

Table 5: Background and Summary of Total Emissions for Each Quarter

<table>
<thead>
<tr>
<th></th>
<th>2006 4th Quarter Total Emissions</th>
<th>2007 2nd Quarter Total Emissions</th>
<th>2008 2nd Quarter Total Emissions</th>
</tr>
</thead>
<tbody>
<tr>
<td>Wind Angle</td>
<td>130</td>
<td>40</td>
<td>150</td>
</tr>
<tr>
<td>Wind Speed (m/s)</td>
<td>0.89</td>
<td>1.34</td>
<td>2.24</td>
</tr>
<tr>
<td>Stability Class</td>
<td>B</td>
<td>B</td>
<td>B</td>
</tr>
<tr>
<td>Background</td>
<td>3.1 ppm</td>
<td>2.45 ppm</td>
<td>1.82 ppm</td>
</tr>
<tr>
<td>Concentration</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Total Emissions</td>
<td>608.7</td>
<td>707.5</td>
<td>1233.4</td>
</tr>
<tr>
<td>(g/s)</td>
<td></td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

The results for total emissions based on the three data sets are quite different as seen above in Table 5. There are various reasons as to why there appears to be an increasing trend seen in Table 5 from year 2006-2008, such as the meteorological history (especially recent
rains), amount of waste dumped prior to the quarterly survey, type of waste dumped, and possible cracks or leaks occurring with an aging system. Another possibility is that the values of the meteorological parameters were not known accurately when calculating the emissions.

**Task 2.1: Sensitivity Study**

The Gaussian equation, used to solve for the methane emission rates, relies heavily on meteorological parameters. Variables such as wind speed, the horizontal spread function, the vertical spread function, are based on the meteorology at that given time. It was important to perform a sensitivity study on these parameters to determine how they affect the total emissions being predicted since there is no hourly on-site meteorological data provided at the landfill. For this study, the 2nd quarter 2007 monitoring report was used and the conditions that were estimated at the time of this survey were used as the initial conditions (Table 6). These values were changed one at a time to determine the sensitivity of each variable in the Gaussian equation.

**Table 6: 2007 2nd Quarter Initial Conditions Used**

<table>
<thead>
<tr>
<th>2007 2nd Quarter</th>
<th></th>
</tr>
</thead>
<tbody>
<tr>
<td>Wind Angle=</td>
<td>40</td>
</tr>
<tr>
<td>Wind Speed=</td>
<td>1.34</td>
</tr>
<tr>
<td>Stability Class=</td>
<td>B</td>
</tr>
<tr>
<td>Background Concentration=</td>
<td>2.45 ppm</td>
</tr>
<tr>
<td>Total Emissions (g/s)=</td>
<td>707.5</td>
</tr>
</tbody>
</table>

The sensitivity study was done on these parameters and the results can be seen in Table 7 below. The table shows how much the total emissions of the landfill are affected by each of the parameters especially stability class. The wind speed and wind angle correlate well when comparing percent change from base and percent change in predicted emissions, while stability
classes percent change from base and predicted emissions vary from class to class. That is why it is crucial to get precise on-site measurements of meteorology to get accurate predicted emissions.

Table 7: Sensitivity Study of 2007 2nd Quarter Meteorological Conditions

<table>
<thead>
<tr>
<th>Wind Speed (u) (m/s):</th>
<th>% Change from Base</th>
<th>Total Emissions (g/s)</th>
<th>% Change in Predicted Emissions</th>
</tr>
</thead>
<tbody>
<tr>
<td>1</td>
<td>-25.4%</td>
<td>528.0</td>
<td>-25.4%</td>
</tr>
<tr>
<td>1.25</td>
<td>-6.7%</td>
<td>660.0</td>
<td>-6.7%</td>
</tr>
<tr>
<td>1.34 (base)</td>
<td>0.0%</td>
<td>707.5</td>
<td>0.0%</td>
</tr>
<tr>
<td>1.5</td>
<td>11.9%</td>
<td>791.9</td>
<td>11.9%</td>
</tr>
<tr>
<td>2</td>
<td>49.3%</td>
<td>1055.9</td>
<td>49.3%</td>
</tr>
<tr>
<td>2.5</td>
<td>86.6%</td>
<td>1319.9</td>
<td>86.6%</td>
</tr>
<tr>
<td>Stability Class:</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>D</td>
<td>-33.3%</td>
<td>281.3</td>
<td>-60.2%</td>
</tr>
<tr>
<td>C</td>
<td>-16.7%</td>
<td>415.1</td>
<td>-41.3%</td>
</tr>
<tr>
<td>B (base)</td>
<td>0.0%</td>
<td>707.5</td>
<td>0.0%</td>
</tr>
<tr>
<td>A</td>
<td>16.7%</td>
<td>1302.7</td>
<td>84.1%</td>
</tr>
<tr>
<td>Theta (θ) (Direction wind from)</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>20</td>
<td>11.11%</td>
<td>820.9</td>
<td>16.0%</td>
</tr>
<tr>
<td>40 (base)</td>
<td>0.00%</td>
<td>707.5</td>
<td>0.0%</td>
</tr>
<tr>
<td>60</td>
<td>11.11%</td>
<td>794.8</td>
<td>12.3%</td>
</tr>
<tr>
<td>320</td>
<td>44.44%</td>
<td>978.2</td>
<td>38.3%</td>
</tr>
</tbody>
</table>

**Task 3: Derive H₂S to Methane Ratio**

The optimal way to derive the H₂S to methane ratio would be to actually sample the H₂S content at various points during the walking survey of the monitoring report, using a portable H₂S analyzer that can measure H₂S to levels as close as 1 ppb. So if both the H₂S and methane concentrations were sampled at various points during the walking survey, a robust sample of the two could be used to provide an accurate H₂S to methane ratio for that particular landfill. The same theory could be applied to any odorous gas at a landfill to derive a useable ratio.
This method was not able to be done for this study, so further research of the literature was done to find reported \( \text{H}_2\text{S}\)-to-methane ratios.

### Table 8: \( \text{H}_2\text{S}/ \text{CH}_4 \) Ratios from Literature Review

<table>
<thead>
<tr>
<th>Case</th>
<th>Hydrogen Sulfide</th>
<th>Methane</th>
<th>( \text{H}_2\text{S}/ \text{CH}_4 ) Ratio</th>
</tr>
</thead>
<tbody>
<tr>
<td>122</td>
<td>63.3 ppm</td>
<td>540000 ppm</td>
<td>0.00012</td>
</tr>
<tr>
<td>223</td>
<td>70.0 ppm</td>
<td>500000 ppm</td>
<td>0.00014</td>
</tr>
<tr>
<td>324</td>
<td>900 ppm</td>
<td>730000 ppm</td>
<td>0.00123</td>
</tr>
<tr>
<td>425</td>
<td>20 ppm</td>
<td>532830 ppm</td>
<td>0.00004</td>
</tr>
<tr>
<td>58</td>
<td>10000 ppm</td>
<td>450000 ppm</td>
<td>0.02222</td>
</tr>
<tr>
<td>627</td>
<td>247.8 ppm</td>
<td>286000 ppm</td>
<td>0.00087</td>
</tr>
<tr>
<td>727</td>
<td>115.3 ppm</td>
<td>585000 ppm</td>
<td>0.00020</td>
</tr>
<tr>
<td>827</td>
<td>2344 ppm</td>
<td>316000 ppm</td>
<td>0.00742</td>
</tr>
</tbody>
</table>

The figures and tables used to determine these ratios are attached in Appendix D. Table 8 has eight different cases that have a wide range of values for the \( \text{H}_2\text{S}/ \text{CH}_4 \) Ratio. Each case was based on a total of 1 million ppm, if all the components of landfill gases were shown (Methane, \( \text{CO}_2 \), Nitrogen, Oxygen, \( \text{H}_2\text{S} \), and other trace chemicals). Cases 1, 2, 4, 6, and 7 are a good representation of how much hydrogen sulfide can be found in an average landfill. Cases 3, 5, and 8 are worst case scenarios that show what the ratio was under extreme circumstances that normally are not seen, but are possible. Case 5 considered \( \text{H}_2\text{S} \) was 1% and accounted for all the trace chemical species, which is an overly conservative approach based on what is known about landfill gas and what was found in the literature review.
In Figueroa’s thesis, Case 4 was the ratio used to determine the odor buffer distances. This ratio was determined to be not conservative enough since literature review of the composition of landfill gases showed all the other ratios to be higher than Case 4. It was determined for this study that a H$_2$S/CH$_4$ ratio of 0.00013 was a more conservative approach and could be an accurate approach based on Cases 1 and 2.

**Task 4: Set Limits**

The task of setting limits was researched in the literature to decide the most reasonable 1-hour H$_2$S limits. Again these limits are for use when dealing with H$_2$S, since different odors have different thresholds. The 1-hour H$_2$S limits of >30 ppb, 15-30 ppb, and under 15 ppb were selected as discussed below: The >30 ppb has been used previously to set an exceedance level for H$_2$S. There are two examples of this. One would be the control of odorous gas at Massachusetts landfills, that is based on an average concentration of hydrogen sulfide measured in the ambient air at a location must be less than or equal to 15 ppb averaged over 8 hours or 30 ppb averaged over one hour. The second example is the California ambient air quality standard (CAAQS) one-hour limit for hydrogen sulfide of 30 ppb. Using these two examples and the researcher’s judgment, the 1-hour H$_2$S limits were selected. Using Equation 1 from the methodology, the H$_2$S limits corresponding to the methane concentrations are listed in Table 9.

**Table 9: Methane Concentration Limits**

<table>
<thead>
<tr>
<th>H$_2$S Limit</th>
<th>Methane Concentration</th>
</tr>
</thead>
<tbody>
<tr>
<td>&lt;15 ppb</td>
<td>&lt;115 ppm</td>
</tr>
<tr>
<td>15-30 ppb</td>
<td>115-231 ppm</td>
</tr>
<tr>
<td>&gt;30 ppb</td>
<td>&gt;231 ppm</td>
</tr>
</tbody>
</table>
Tables similar to this can be created for other odors if it is determined H₂S is not the main contributor to odors at a certain landfill. Making the odor limits into methane concentrations also helps when generating plots because instead of making multiple plots for different odors anyone can just pull up the methane concentration limits for that odor and read them off one plot of methane concentrations for that landfill.

**Task 5: Air Dispersion Modeling**

The same methodology discussed previously was used to determine the methane concentration around the SCL landfill using the air dispersion modeling program AERMOD.

The meteorological data gathered is from January 1, 1999 to December 31, 2003. The hourly surface data is from the Orlando International Airport and the upper air data is from Ruskin, FL. The meteorological data for the year of 2001 was used for the research of the SCL, but any of the four years can be used. A 7.5-minute horizontal datum DEM file for Osceola was used creating two output files, one for the sources and one for the receptors. The receptor file was used when running AERMOD since the terrain is relatively the same as when the DEM file was created, but the source file was not used since we are dealing with a landfill whose terrain is constantly changing. Instead the topographic maps provided with the Surface Emission Monitoring reports were used to get accurate measurements of the terrain inside the landfill boundaries.

The input file was set up with an averaging time of 1-hour and the pollutant modeled in this case was methane. The source pathway specified the amount of sources, the emission rate for each source, and the base elevation of each source. The 4th Quarter 2006 had 178 point sources, the 2nd Quarter 2007 had 108 point sources, and the 2nd Quarter 2008 had 170 point
sources (all these point sources were determined from task 2). AERMOD was then run (with 2001 meteorological data) and the 4\textsuperscript{th} highest concentrations of methane in (ug/m\textsuperscript{3}) were recorded at all 300 receptors put in excel and converted to parts per million (ppm). The receptors 1-252 are the polar rings (36 receptors on each ring every 10 degrees) which correspond to distances of 800m, 900m, 1000m 1100m 1200m, 1300m, and 1400m from the center of the landfill. The receptors 252-300 represent the fence line of the Seminole County Landfill.

The results are presented in Appendix E for the 4\textsuperscript{th} quarter 2006, 2\textsuperscript{nd} quarter 2007, and 2\textsuperscript{nd} quarter 2008. It should be noted that most of the concentrations seen in Appendix E occurred during the hours of 2300-0900 (or 11p.m.-9a.m.). This is discussed further in the recommendations section and could lead to modeling for hours when most nuisance complaints happen rather then a full year of meteorological data, based on input from the community.
**Task 6: Generate Plots**

Using the methodology that was presented earlier, the results from the air dispersion modeling were used to generate color coded plots.

Since \( \text{H}_2\text{S} \) was used in this study the methane concentration limits for \( \text{H}_2\text{S} \) from Table 9 were used to create the plots for each quarter. \( \text{H}_2\text{S} \) is not the only odorous compound found in landfills and the ratio of an odorous compound to methane will change so Table 9 is only relevant for this study of \( \text{H}_2\text{S} \).

The three limits are color coordinated with the colors of red, yellow, and green. The red zone is a “do not build zone” and corresponds to \( \text{H}_2\text{S} \) concentrations > 30ppb (or methane > 231 ppm). The yellow zone is a “proceed with caution” zone and corresponds to a \( \text{H}_2\text{S} \) from 15-30ppb (or methane from 115-231 ppm). The green zone is a “safe to build zone” and corresponds to \( \text{H}_2\text{S} \) concentrations < 15ppb (or methane < 115 ppm).

The three-color plots in the following pages are the findings from the 4\(^{\text{th}}\) Quarter 2006, 2\(^{\text{nd}}\) Quarter 2007, and 2\(^{\text{nd}}\) Quarter 2008 in order. From the three quarters of data the plots are expected to show an increasing trend of concentrations of \( \text{H}_2\text{S} \), based primarily on the apparent increasing trend of total emissions from the landfill from 2006-2008.

The part that is not easily determined is where these expected higher concentrations are going to occur outside the landfill fence line from quarter to quarter. The meteorology experienced during a certain year plays a big role in this, but what was looked at before doing the air dispersion modeling is where the greatest point source emissions were coming from in the landfill. The isopleths of the point source methane emissions in grams/sec (created from the MATLAB results) are shown by quarter directly before the color generated plot of the air dispersion modeling figures. These isopleths of point source methane emissions give some clues
to where the greatest concentrations of odors may occur even before air dispersion modeling is done.

Based on Figure 12 it was expected that the highest concentrations of odors are just outside the northeast part of the landfill, since this was where the greatest amounts of methane were being emitted. After doing the air dispersion modeling and looking at Figure 13 it can be inferred that the high methane emissions in Figure 12 are what caused the high concentrations of methane/odor just outside the northern fence line.
Figure 13: 4th Quarter 2006 Concentration Limit Plot

Figure 14 showing the 2nd Quarter 2007 methane emission isopleths was also a good indicator of where the highest concentrations of methane/odor were going to occur. This time it was expected the highest concentrations of methane/odor were going to occur around the north-east quadrant of the landfill, with a smaller localized high just south of the landfill.
Figure 14: 2\textsuperscript{nd} Quarter 2007 Methane Emission Isopleths
Figure 15 proves once again that the methane emission isopleths were a good indicator of where the high methane/odor concentrations would occur. Although the methane emissions isopleths indicated higher concentrations in the northeast and the higher concentration just to the south, the emission isopleths did not give a good indication of the higher concentration along the western edge of the landfill. Air dispersion modeling was the only way to pin point where the highest concentrations of methane/odors occurred in this case.
Figure 16: 2\textsuperscript{nd} Quarter 2008 Methane Emission Isopleths

The 2\textsuperscript{nd} Quarter 2008 data is a great example of how odors around the landfill and methane emissions within the landfill are changing. Looking at Figure 17 we had an extremely low chance of odors occurring to the north and south of the landfill, while a high chance of high concentrations of methane/odors to the west and even higher methane/odor concentrations to the east existed. Modeling numerous quarters of data gives a greater level of confidence for determining where odors are occurring at any given time.
Figure 17: 2nd Quarter 2008 Concentration Limit Plot
CHAPTER 5: CONCLUSIONS

Being able to predict where, at what concentration, and when odors are occurring is an issue that is affecting people moving into houses around landfills. So as odor complaints increase, the managing of odors is becoming a bigger issue. Using this thesis as the foundation and framework to determine odor buffer distances will help in increasing effective odor management. This thesis gives people the opportunity to evaluate how close they want to build to landfills and what they might experience in the way of odors depending on how close they do build.

This thesis also presents a new and promising way of estimating methane emissions from landfills. By using the Gaussian equation and the inverse modeling method, the total emissions can be estimated from a landfill and the point sources that are emitting methane within the landfill can estimated. The conclusions of this research include:

1) A methodology for estimating methane emissions from a landfill using the surface emissions monitoring reports taken every quarter has been developed. Using hundreds of ambient VOC measurements, and an inverse modeling method, the methane emissions can be calculated, both on a total basis and with a spatial distribution.

2) A methodology to estimate and derive an odorous gas to methane ratio was proposed. This ratio can be used to set limits that will help create odor buffer distances.

3) The air dispersion model AERMOD can be used to determine where these odors will occur and at what concentration.
4) Using the results from AERMOD and limits set for an odorous gas colored plots can be created to establish odor buffer distances.
CHAPTER 6: RECOMMENDATIONS

Further study is recommended to produce an even more accurate/specific methodology in estimating methane emissions and determining odor buffer distances. To accomplish this there are a series of further steps and analyses that can be done.

First the H$_2$S to methane ratio or any other odorous gas to methane ratio should be measured at the landfill during the required quarterly surface emissions monitoring report, rather then using average ratios from literature review. Also during the quarterly surface emission monitoring more accurate measurements of the meteorological parameters need to be done on site, since the estimation of methane emissions is so sensitive to these parameters. It is also recommended that several samples outside the landfill be tested for concentrations of the odorous gas to compare with the air dispersion modeling results. This will help in giving more accurate estimations of the odor buffer distances.

The next recommendation is to perform analysis of more quarters worth of data and to model another landfill. Using more quarters would not only give a more robust data set to work with but also determine if odors dramatically change from season to season (as of right now only two 2$^{nd}$ Quarters and one 4$^{th}$ quarter worth of emissions have been used). Using multiple landfills would also test the sensitivity of a landfill to the heights of a landfill, the surroundings of a landfill, the odorous gas to methane ratio (if one can be sampled), and the emissions from the landfill.

Future studies should include a more in-depth look at when odors occur and if that can affect the odor buffer distances. Most of the odors in this research were occurring at night when people are sleeping. This suggests research should concentrate more on odors that occur during the morning, afternoon, and early evening when more people are influenced by these odors.
Doing multiple landfills and quarters of data would help in determining if the highest concentrations of odors occur at night when most people are sleeping.
APPENDIX A: SEMINOLE COUNTY LANDFILL VOC MEASUREMENTS (CONDUCTED BY SCS FIELD SERVICES)
January 30, 2007

SCS File No. 09206066.02

David Gregory
Solid Waste Director
Seminole County Landfill
1930 E. Osceola Road
Geneva, FL 32733-7499

Subject: Results of Fourth Quarter 2006 Surface Emissions Monitoring, Seminole County Landfill, Geneva, FL

Dear David:

On December 22, 2006, SCS Field Services (SCS-FS) conducted surface emissions monitoring at the subject location, as specified in 40 CFR 60.755 (c) and (d), and 40 CFR 60, Appendix A, Method 21. These regulations require surface monitoring around the perimeter of and within the LFG Collection System area, the extent of which includes landfill areas where waste exceeds two years in age at final grade or five years in age at interim grades.

Monitoring was completed in accordance with the regulations. A total of 425 points were surveyed for emissions of volatile organic compounds (VOC), as methane, using a Landec SEM 500 flame ionization detector.

Over the area surveyed, no points exceeded 500 ppm above background. As such, no additional monitoring or remedial action was required for this quarter. The monitoring data is attached.

SCS-FS, an employee-owned company, thanks you for using our services. Please contact either of the undersigned if you require further information.

Sincerely,

[Signature]
Garold (Tony) A. Cartee
Project Manager
SCS FIELD SERVICES

Peter J. Carrico
Vice President
SCS FIELD SERVICES

Attachment
SCS Field Services, Inc.

NSPS Surface Emissions Monitoring
Calibration and Pertinent Data Form

Date: 22 Dec 96 Site Name: Seminole County LF Job Number: 09206066

Technician(s): Scott Lambert

Weather Observations

Wind Speed: 0.2 MPH Wind Direction: 36° SE Barometric Pressure: 30.12 "Hg

Air Temperature 68°F General Weather Conditions: Clear

Calibration Information
Pre-monitoring Calibration Precision Check

Procedure: Calibrate the instrument. Make a total of three measurements by alternating zero air and the calibration gas. Record the readings and calculate the average algebraic difference between the instrument reading and the calibration gas as a percentage. The calibration precision must be less than or equal to 10% of the calibration gas value.

Instrument ID: TDA Cal Gas Concentration: 500 ppm

<table>
<thead>
<tr>
<th>Trial</th>
<th>Zero Air Reading</th>
<th>Cal Gas Reading</th>
<th>Cal Gas Conc. – Cal Gas Reading</th>
</tr>
</thead>
<tbody>
<tr>
<td>1</td>
<td>0.94</td>
<td>504</td>
<td>500</td>
</tr>
<tr>
<td>2</td>
<td>0.98</td>
<td>500</td>
<td>500</td>
</tr>
<tr>
<td>3</td>
<td>0.97</td>
<td>501</td>
<td>501</td>
</tr>
</tbody>
</table>

Average Difference: 1.6

Calibration Precision = Average Difference/Cal Gas Conc. X 100%

= 1.6/500 X 100%

= 0.32 %

Post-monitoring Calibration Check

Zero Air Reading: 0.98 ppm Cal Gas Reading 505 ppm

Background Concentration Checks

Upwind Location Description: SE Side of LF Reading: 3.10 ppm
Downwind Location Description: NW Side of LF Reading: 6.12 ppm

Notes/Comments

No exceedances. High reading in area of 246 thru 252; dry brush cover.
<table>
<thead>
<tr>
<th>Date</th>
<th>Time</th>
<th>Monitoring Tag</th>
<th>FID Concentration</th>
<th>Unit</th>
<th>Notes</th>
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</thead>
<tbody>
<tr>
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<td>7:47:07</td>
<td>500PPM</td>
<td>504</td>
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</tr>
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<td>22-Dec-06</td>
<td>7:49:03</td>
<td>ZERO GAS</td>
<td>0.96</td>
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<td>UPWIND</td>
<td>3.1</td>
<td>PPM</td>
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<td>7:55:18</td>
<td>DOWN WIND</td>
<td>6.12</td>
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</tr>
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<td>8:02:04</td>
<td>1</td>
<td>3.89</td>
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<td>22-Dec-06</td>
<td>8:02:30</td>
<td>2</td>
<td>1.79</td>
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<td>22-Dec-06</td>
<td>8:02:50</td>
<td>3</td>
<td>1.87</td>
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<td>OK</td>
</tr>
<tr>
<td>22-Dec-06</td>
<td>8:03:09</td>
<td>4</td>
<td>2.23</td>
<td>PPM</td>
<td>OK</td>
</tr>
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<td>22-Dec-06</td>
<td>8:03:28</td>
<td>5</td>
<td>2.42</td>
<td>PPM</td>
<td>OK</td>
</tr>
<tr>
<td>22-Dec-06</td>
<td>8:03:41</td>
<td>6</td>
<td>2.59</td>
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</tr>
<tr>
<td>22-Dec-06</td>
<td>8:04:00</td>
<td>7</td>
<td>1.81</td>
<td>PPM</td>
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<tr>
<td>22-Dec-06</td>
<td>8:04:19</td>
<td>8</td>
<td>1.62</td>
<td>PPM</td>
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<tr>
<td>22-Dec-06</td>
<td>8:04:39</td>
<td>9</td>
<td>1.83</td>
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<td>22-Dec-06</td>
<td>8:04:59</td>
<td>10</td>
<td>1.81</td>
<td>PPM</td>
<td>OK</td>
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<tr>
<td>22-Dec-06</td>
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<td>11</td>
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<td>2.14</td>
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<td>3.2</td>
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<td>15</td>
<td>5.12</td>
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<td>OK</td>
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<td>5.14</td>
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<tr>
<td>22-Dec-06</td>
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<td>17</td>
<td>26.33</td>
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<td>OK</td>
</tr>
<tr>
<td>22-Dec-06</td>
<td>8:07:50</td>
<td>18</td>
<td>25.83</td>
<td>PPM</td>
<td>OK</td>
</tr>
<tr>
<td>22-Dec-06</td>
<td>8:08:10</td>
<td>19</td>
<td>11.42</td>
<td>PPM</td>
<td>OK</td>
</tr>
<tr>
<td>22-Dec-06</td>
<td>8:08:33</td>
<td>20</td>
<td>12.81</td>
<td>PPM</td>
<td>OK</td>
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<tr>
<td>22-Dec-06</td>
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<td>22-Dec-06</td>
<td>8:09:23</td>
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</tr>
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<td>23</td>
<td>12.31</td>
<td>PPM</td>
<td>OK</td>
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<tr>
<td>22-Dec-06</td>
<td>8:10:07</td>
<td>24</td>
<td>83.95</td>
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<tr>
<td>22-Dec-06</td>
<td>8:12:20</td>
<td>25</td>
<td>26.83</td>
<td>PPM</td>
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</tr>
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4th Quarter 2006 SEM Monitoring Data

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History for Sanford, FL
Friday, December 22, 2006

Daily Summary

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Moisture:
- Dew Point: 61 °F / 16 °C
- Average Humidity: 78
- Maximum Humidity: 94
- Minimum Humidity: 56

Precipitation:
- Precipitation: 2.15 in / 5.46 cm

Sea Level Pressure:
- Sea Level Pressure: 30.09 in / 1019 hPa

Wind:
- Wind Speed: 3 mph / 5 km/h (South)
- Max Wind Speed: 18 mph / 29 km/h
- Max Gust Speed: 26 mph / 45 km/h
- Visibility: 8 miles / 13 kilometers
- Events: Fog, Rain, Thunderstorm

Averages and records for this station are not official NWS values.
Click here for data from the nearest station with official NWS data (KMCO).
T = Trace of Precipitation, MM = Missing Value

Source: NWS Daily Summary
## Hourly Observations

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Figure A-1: Seminole County Landfill 4th Quarter 2006 VOC Readings
July 10, 2007
SCS File No. 09206066.02

David Gregory
Solid Waste Director
Seminoe County Landfill
1930 E. Osceola Road
Geneva, FL 32773-7499

Subject: Results of Second Quarter 2007 Surface Emissions Monitoring, Seminole County Landfill, Geneva, FL

Dear David:

On June 29, 2007, SCS Field Services (SCS-FS) conducted surface emissions monitoring at the subject location, as specified in 40 CFR 60.755 (c) and (d), and 40 CFR 60, Appendix A, Method 21. These regulations require surface monitoring around the perimeter of and within the LFG Collection System area, the extent of which includes landfill areas where waste exceeds two (2) years in age at final grade or five years in age at interim grades.

Monitoring was completed in accordance with the regulations. A total of 358 points were surveyed for emissions of volatile organic compounds (VOC), as methane, using a Foxboro TVA-1000B flame ionization detector.

Over the area surveyed, no points exceeded 500 ppm above background. As such, no additional monitoring or remedial action was required for this quarter. The monitoring data is attached.

SCS-FS, an employee-owned company, thanks you for using our services. Please contact either of the undersigned if you require further information.

Sincerely,

Michael D. Knox
Project Manager
SCS FIELD SERVICES

Initial of Author/Initials of Reviewer:

Attachments: Calibration Form
Monitoring Data Table
SCS Field Services, Inc.

NSPS Surface Emissions Monitoring
Calibration and Pertinent Data Form

Date: 04/29/07  Site Name: SEMINOLE LE  Job Number: 868766.07

Technician(s): Jason Bluer

Weather Observations

Wind Speed: 3 MPH  Wind Direction: NE  Barometric Pressure: 30.01 "Hg

Air Temperature: 88°F  General Weather Conditions: Cloudy

Calibration Information

Pre-monitoring Calibration Precision Check

Procedure: Calibrate the instrument. Make a total of three measurements by alternating zero air and the calibration gas. Record the readings and calculate the average algebraic difference between the instrument reading and the calibration gas as a percentage. The calibration precision must be less than or equal to 10% of the calibration gas value.

Instrument ID: TVA2000  Cal Gas Concentration: 500 ppm

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Average Difference: 3

Calibration Precision = Average Difference/Cal Gas Conc. X 100%

= 3/500 X 100%

= 0.6%  

Post-monitoring Calibration Check

Zero Air Reading: 1.87 ppm  Cal Gas Reading 53 ppm

Background Concentration Checks

Upwind Location Description: Road NE from Reading: 2.45 ppm
Downwind Location Description: Road SW across Reading: 1.87 ppm

Notes/Comments

No Exceedances
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NSPS SURFACE EMISSIONS MONITORING DATA
SECOND QUARTER 2007

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### TABLE I. SEMINOLE COUNTY LANDFILL
NSPS SURFACE EMISSIONS MONITORING DATA
SECOND QUARTER 2007

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<th>Unit</th>
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History for Sanford, FL
Friday, June 29, 2007

Daily Summary

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<td>98 °F / 36 °C (1959)</td>
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Moisture:

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<td>Precipitation</td>
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Sea Level Pressure:

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<td>Sea Level Pressure</td>
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Wind:

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<tr>
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<tr>
<td>Max Gust Speed</td>
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<tr>
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<tr>
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Averages and records for this station are not official NWS values.

Click here for data from the nearest station with official NWS data (KMCQ).

T = Trace of Precipitation, MM = Missing Value
Source: NWS Daily Summary

Seasonal Weather Averages
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<th>Sea Level Pressure:</th>
<th>Visibility:</th>
<th>Wind Dir</th>
<th>Wind Speed:</th>
<th>Gust Speed:</th>
<th>Precip:</th>
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<th>Conditions:</th>
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<td>10.0 miles</td>
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<td>N/A</td>
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<td>69.1°F</td>
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<td>10.0 miles</td>
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<td>10.0 miles</td>
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<td>68.1°F</td>
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<td>10.0 miles</td>
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<td>30.04 in./hPa</td>
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<td>69.1°F</td>
<td>30.05 in./hPa</td>
<td>10.0 miles</td>
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<td>69.8°F</td>
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<td>71.1°F</td>
<td>30.05 in./hPa</td>
<td>10.0 miles</td>
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<td>71.6°F</td>
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<td>72.0°F</td>
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<td>30.03 in./hPa</td>
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<td>Mostly Cloudy</td>
<td></td>
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<tr>
<td>3:24 PM</td>
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<td>72.8°F</td>
<td>30.03 in./hPa</td>
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<tr>
<td>3:41 PM</td>
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<td>69.8°F</td>
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<td>N/A</td>
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<td>Scattered Clouds</td>
<td></td>
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<tr>
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<td>69.1°F</td>
<td>30.01 in./hPa</td>
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<td>N/A</td>
<td>Scattered Clouds</td>
<td></td>
</tr>
<tr>
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<td>1016.3 hPa</td>
<td>16.1 kilometers</td>
<td>4.5 mph</td>
<td>4.5 mph</td>
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<td>N/A</td>
<td>Scattered Clouds</td>
<td></td>
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<tr>
<td>5:53 PM</td>
<td>30.6°C</td>
<td>20.6°C</td>
<td>1016.3 hPa</td>
<td>16.1 kilometers</td>
<td>4.5 mph</td>
<td>4.5 mph</td>
<td>N/A</td>
<td>N/A</td>
<td>Scattered Clouds</td>
<td></td>
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</tbody>
</table>
Figure A-2: Seminole County Landfill 2nd Quarter 2007 VOC Reading Locations
July 10, 2008  
SCS File No. 09206066.09

Johnny Edwards  
Solid Waste Manager  
Seminole County Landfill  
1930 E. Osceola Road  
Geneva, FL 32773-7499

Subject: Results of Second Quarter 2008 Surface Emissions Monitoring, Seminole County Landfill, Geneva, FL

Dear Johnny:

On June 26, 2008, SCS Field Services (SCS-FS) conducted surface emissions monitoring (SEM) at the subject location, as specified in 40 CFR 60.755 (c) and (d), and 40 CFR 60, Appendix A, Method 21.

Monitoring was completed in accordance with the regulations. A total of 442 points were surveyed for emissions of volatile organic compounds (VOC), as methane, using a Foxboro TVA-1000B flame ionization detector.

The monitoring data are presented in Table 1. The monitoring route was determined in the field by SCS-FS. The Calibration and Pertinent Data Forms are also attached. During monitoring the TVA-1000 did not record all of the monitoring points. There is no logged information for readings 222 thru 244. The readings for these locations were manually observed as being below 500 ppm. All points with readings 500 ppm above background were recorded on the route map. There were no readings that were recorded at 500 ppm above background for locations 222 thru 244.

OBSERVATIONS
Over the area surveyed, five monitoring points (Tag numbers 38, 61, 73, 77, 212) had emissions above 500 ppm VOCs. Remedial action was done by increasing vacuum to the wells that were closest to the exceedance areas. Each of these exceedances passed the 10-day recheck with readings below 500 ppm. Per the regulations, these locations will need to be rechecked one month from June 26, 2008, the date of the original exceedances.
SCS-FS, an employee-owned company, thanks you for using our services. Please contact either of the undersigned if you require further information.

Sincerely,

Jason Bever       Michael D. Knox
Senior Field Technician     Project Manager
SCS FIELD SERVICES     SCS FIELD SERVICES

Attachments:  Calibration Form
               Monitoring Data Table
### TABLE 1. SEMINOLE COUNTY LANDFILL
NSPS SURFACE EMISSIONS MONITORING DATA
SECOND QUARTER 2008

<table>
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<tr>
<th>Date</th>
<th>Time (Hr:Min:Sec)</th>
<th>Monitoring Tag</th>
<th>FID Concentration</th>
<th>Unit</th>
<th>Notes</th>
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<td>502</td>
<td>PPM</td>
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NA = No data recorded
### History for Sanford, FL

**Thursday, June 26, 2008**

#### Daily Summary

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Averages and records for this station are not official NWS values. Click here for data from the nearest station with official NWS data (KMCQ).

T = Trace of Precipitation, MM = Missing Value

**Source:** NWS Daily Summary

[Seasonal Weather Averages]
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<th>Sea Level Pressure</th>
<th>Visibility</th>
<th>Wind Dir</th>
<th>Wind Speed</th>
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<th>Precip Events</th>
<th>Conditions</th>
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Temperature and wind conditions for different times and days are listed, along with visibility and weather conditions.
Figure A-3: Seminole County Landfill 2nd Quarter 2008 VOC Readings
APPENDIX B: ESTIMATING METHANE EMISSIONS (INPUT-CONCENTRATION FILES)
(The X and Y coordinates are in Universal Transverse Mercator (UTM) which is the metric equivalent of latitude and longitude in map projections. The last column C is the concentration of methane recorded at that receptor in units of \( \text{ug/m}^3 \) during the surface emissions monitoring report minus background concentrations.)

### 4th Quarter 2006

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MATLAB Predicted Methane Emissions 2nd Quarter 2007:

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MATLAB Predicted Methane Emissions 2nd Quarter 2008:

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APPENDIX D: AVERAGE LANDFILL BIOGAS COMPOSITION
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<tr>
<td>Carbon dioxide, CO₂</td>
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<tr>
<td>Nitrogen, N₂</td>
</tr>
<tr>
<td>Oxygen, O₂</td>
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<tr>
<td>Chlorine</td>
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<td>Fluorine</td>
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Case 2:

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<td>Methane</td>
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<td>Carbon dioxide</td>
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<tr>
<td>Nitrogen</td>
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<td>Heavier hydrocarbons</td>
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Case 3:

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**Case 4:**

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**Trace Gases**

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Case 5:

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Case 6:

At three landfill sites in South Korea, samples of LFG vents yielded following concentrations of H$_2$S in ppb

Site W-247,758 ppb (average of 6 samples, with readings ranging from 212 to 681,370 ppb, with 3 samples above 225,000). Methane readings averaged: 28.6%

Site B- 115,275 ppb (average of 8 samples, with readings ranging from 89,132 to 143,091 ppb, with only 1 sample below 102,000). Methane reading averaged 58.5%

Site H- 2,344,360 ppb (average of 3 samples: 854,580 ppb, 5,142900 ppb, and 1,035,600 ppb). Methane readings averaged 31.6%

APPENDIX E: 4TH RANK AERMOD OUTPUTS
Shown below are the outputs from the air dispersion model AERMOD each of the concentrations at a given receptor are concentrations of methane that can be used as a surrogate for odors. Each of the receptors is either part of one of the polar rings that surround the landfill or is part of the fence line receptors.

4th Quarter 2006

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<th>Y (meters)</th>
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<th>Concentration (ppm)</th>
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